

1 **Examination of uncertainty in per unit area estimates of**
2 **aboveground biomass using terrestrial lidar and ground**
3 **data.**

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15 **Abstract:** In estimating aboveground forest biomass (AGB), three sources of
16 error that interact and propagate include: (1) measurement error, the quality of the
17 tree-level measurement data used as inputs for the individual-tree equations; (2)
18 model error, the uncertainty about the equations of the individual trees; and (3)
19 sampling error, the uncertainty due to having obtained a probabilistic or
20 purposive sample, rather than a census, of the trees on a given area of forest land.
21 Monte Carlo simulations were used to examine measurement, model and
22 sampling error, and to compare total uncertainty between models, and between a
23 phase-based terrestrial laser scanner (TLS) and traditional forest inventory
24 instruments. Input variables for the equations were diameter at breast height, total
25 tree height (defined the height from the uphill side of the tree to the tree top) and
26 height to crown base; these were extracted from the terrestrial LiDAR data.

27 Relative contributions for measurement, model and sampling error were 5%, 70%
28 and 25%, respectively when using TLS, and 11%, 66% and 23%, respectively
29 when using the traditional inventory measurements as inputs into the models. We
30 conclude that the use of TLS can reduce measurement errors of AGB compared
31 to traditional measurement approaches.

32 **Keywords:** Model error; sampling error; measurement error; Pacific Northwest
33

34 **1. Introduction**

35 Forest inventory and monitoring programs such as the United States Department of
36 Agriculture (USDA) Forest Inventory and Analysis Program (FIA) produce estimates and
37 reports of forest resources that bear increasing utility for agencies and other users alike.
38 Inventory attributes derived from such estimates often lack a defensible magnitude of
39 certainty to support forest management decisions that satisfy an array of ecological, economic
40 and social requirements. An accurate depiction of the precision of such estimates would serve
41 to guide and support such decisions. With the growing use of FIA inventory data for
42 attributes such as aboveground biomass (AGB), gains in precision made by addressing
43 specific sources of uncertainty could benefit forest managers and planners, as well as
44 scientists drawing inference and making decisions from their AGB estimates (Temesgen et al.
45 2015).

46 The reliability of AGB estimates produced using sampling approaches such as FIA depends
47 on three primary sources of uncertainty that interact and propagate: (1) the quality of the tree-
48 level measurements used as inputs for estimating biomass of individual trees; (2) the
49 uncertainty about the models used for predictions; and (3) the uncertainty due to having
50 obtained a probabilistic sample, rather than a census, of the trees on a given area of forest
51 land (Cunia 1965). Increasing emphasis on acquiring highly accurate estimates of AGB for
52 management and policy decision making also requires transparent characterizations of
53 associated uncertainty stemming from the three sources of error mentioned above. Accurate
54 estimation of these uncertainties requires accounting for all three of the aforementioned
55 sources of uncertainty when constructing reliability statements for AGB. However, many
56 forest inventory operations currently only account for sampling uncertainties, as the first two
57 uncertainties listed are more difficult to estimate from field-based measures alone, and are
58 often assumed to be of less importance. Besides allowing more confidence in landscape level

59 predictions, estimation of all three sources of uncertainty may also provide an opportunity to
60 observe possible gains in precision to be had by addressing uncertainty that arises due to
61 issues with tree-level explanatory measurement data. These have practical implications for
62 instance in terms of the choice of instrument, calibration and standardized training and
63 implementation procedures for data collection (Weiskittel et al. 2011, p.277 and Temesgen et
64 al. 2007).

65 Difficulties in estimating accuracy of tree-level measurements include the collection of
66 suitable ground truth data to base uncertainty estimates on. Henning and Radtke (2006)
67 compared diameter outside bark (DOB) measurements of nine destructively sampled loblolly
68 pine (*Pinus taeda*) trees to the same DOB measurements obtained using a terrestrial laser
69 scanner (TLS). DOBs, measured in 1m intervals, were reported to be within 1-2cm, with
70 greater accuracy achieved for stem portions below the base of live crown. Bienert et al.
71 (2006) and Maas et al. (2008) reviewed and compared work flow and data processing
72 procedures for extracting common inventory attributes such as DOBs and total tree height
73 (HT). As an alternative to these destructive methods, TLS may provide new opportunities to
74 provide ground truthing estimates for current inventory approaches, which typically predict
75 tree metrics based upon a few easily acquired measurements, such as diameter at breast
76 height (DBH) and height. TLS may help us to improve upon these estimates by providing
77 high density point clouds, useful for accurately depicting stem properties, including taper, as
78 well as crown metrics, including crown density and leaf area. For instance, TLS has been
79 used for measuring tree-level metrics such as DOBs and bole heights (Simonse et al. 2003,
80 Hopkinson et al. 2004, Henning and Radtke 2006, Bienert et al. 2006, Maas et al. 2008, Weiß
81 2009, Poeschel et al. 2013, Liang et al. 2014) as well as crown metrics such as height to
82 crown base and crown volume (Chasmer et al. 2006, Jung et al. 2011).

83 Hauglin et al., (2013) determined the biomass of Norway spruce with TLS using voxel-based
84 approaches and crown dimension features. Other techniques include stem reconstruction (Yu
85 et al., 2013), as well as total tree reconstruction (Calders et al., 2014 and Hackenberg et al.,
86 2014). The performance of TLS in obtaining specific individual-tree variables has been
87 demonstrated, including taper (Thies et al. 2004), DOB (Simonse et al. 2003, Hopkinson et
88 al. 2004, Henning and Radtke 2006, Bienert et al. 2006, Maas et al. 2008, Weiß 2009,
89 Pueschel et al. 2013), canopy metrics such as crown area, crown volume and height to crown
90 base (HTCB) (Chasmer et al. 2006, Jung et al. 2011), and bole reconstruction for stem
91 volume calculation (Yu et al. 2013). Chasmer et al. (2006) used coinciding ALS and TLS
92 data to compare against field-based plot measurements of HT, HTCB and maximum crown
93 width. Average height estimate biases were similar for both ALS and TLS at 1.1m and 1.2m,
94 respectively. ALS overestimated HTCB by an average of 1.4m due to point density
95 distributions being weighted toward the top of the tree, whereas TLS underestimated HTCB
96 by 6.4m, not only resulting from the inverse of the aforementioned distribution due to an
97 inverted perspective, but largely due to not accounting for the occurrence of dead branches.

98 Unique to this study is the depiction of how the measurement performance of TLS in
99 extracting these tree-level variables translates into differences in per unit area estimates of
100 forest-related parameters, specifically AGB. We investigate how the total propagated error of
101 AGB associated with using a TLS compares to that associated with using common forest
102 inventory instruments used for standing tree measurements. To do so, we used data from
103 three types of measurements performed on 25 lodgepole pine (*Pinus contorta* Douglas) trees
104 as the basis for making these comparisons. We validate the estimates obtained from TLS and
105 traditional forest inventory instruments against estimates obtained by destructive sampling
106 methods. Using a newly developed set of Component Ratio Method (CRM) equations for
107 predicting lodgepole pine AGB, a Monte Carlo simulation approach was employed for

108 making comparisons between associated uncertainties of per unit area estimates of AGB for
109 each measurement method.

110 **2. Methods**

111 *2.1. Study locations*

112 In order to capture some regional differences in tree form, the data for this study were
113 collected from both the Willamette National Forest (WNF) and the Deschutes National Forest
114 (DNF) in western and central Oregon, USA, respectively. All locations were within an
115 intermediate-elevation range, with the WNF locations spanning from 1,160-1,340 meters
116 above sea level and the DNF locations from 1,280 to 1,340 meters. The WNF locations
117 encompassed two forest types: (1) a diverse mixed-species coniferous forest, with observed
118 species being Douglas-fir (*Pseudotsuga menziesii* var. *menziesii*), western hemlock (*Tsuga*
119 *heterophylla* (Raf.) Sarg.), lodgepole pine, mountain hemlock (*Tsuga mertensiana* (Bong.)
120 Carr.), noble fir (*Abies procera* Rehder), Engelmann spruce (*Picea engelmannii*
121 Parry ex Engelm.), and western white pine (*Pinus monticola* Douglas ex D. Don); and (2) a
122 homogenous coniferous forest composed of primarily lodgepole pine and with a small
123 element of grand fir (*Abies grandis* (Douglas ex D. Don) Lindley). The DNF locations also
124 included one forest type of homogenous coniferous species composition, with observed
125 species being lodgepole pine and ponderosa pine (*Pinus ponderosa* Douglas ex C.Lawson).

126 *2.2. Field measurement approach*

127 Trees were selected via subjectively and common forest inventory variables including DBH,
128 HT and crown ratio (CR) were recorded. While the requirement for accessibility for felling
129 limited our ability to select trees randomly, efforts were taken to select sample trees either
130 located in different forest stands, or sufficiently distanced apart to avoid issues of spatial

131 autocorrelation. A total of 25 trees were destructively measured over a four week period
132 during July and August 2013. DBH, HT and CR ranged from 13.5 to 42.9 cm, 9.2 to 31.9 m
133 and 0.30 to 0.948, respectively. Standing-tree measurements (STM) were conducted prior to
134 felling, with DBH being measured using a Spencer combination tape and with both HT and
135 HTCBB being measured using a Trupulse Laser Rangefinder 360R. For this study, HTCBB was
136 defined as the bole height of the first live limb (i.e., the lowest branch with green needles on
137 it). Among the measurements taken to obtain reference values of AGB, downed-tree
138 estimates of HT and HTCBB were measured with an open reel fiberglass tape.

139 For estimation of component biomass per unit area, ground plot data were collected from
140 those forest stands from which the 25 sample trees were sourced. This ground plot data
141 consisted of 25 cluster plots, each comprised of four circular fixed area subplots arranged
142 around each sample tree. A 0.017 hectare plot was the primary subplot (radius 7.33 m), with
143 the pith of the sample tree as the center. The centers of the other three circular subplots were
144 located 36.58 m at azimuths of 120, 240 and 360° from the pith of the sample tree. The area
145 of these other three subplots was 0.008 hectares (radius 5.18 m). Within each subplot, all
146 trees (> 10.16 cm diameter) were measured and/or recorded for attributes such as species,
147 DBH, HT and HTCBB, among others.

148

149 *2.3. TLS Field Scanning Protocol*

150 In addition to the standing tree measurements, sample trees were scanned with a tripod-
151 mounted FARO Focus^{3D} 120 TLS prior to felling. As opposed to the more common time-of-
152 flight TLS technology the FARO scanner uses phase shift technology which uses the shifts of
153 modulated waves of returned infrared light pulses to calculate distances traveled (FARO
154 2014). Maximum ranges of phase-based scanners are less than those of time-of-flight

155 scanners; however, measurement rates (pulses per second emitted) are usually much higher
156 with greater distance accuracies realized than for time-of-flight scanners. See Table 1 for the
157 technical data of the FARO Focus^{3D} 120.

158 Each sample tree was scanned from three locations around its periphery at distances ranging
159 from approximately 3-8m away from the tree. Scan positions were placed at 120° apart from
160 each other to maximize the information gathered for characterizing the geometric shape of
161 the tree. For automatic co-registration, four manually placed targets were positioned near the
162 sample tree with a minimum of three targets being visible from each scan position. Target
163 construction consisted of printed checkerboard signs affixed to wooden staked panel boards.
164 Because it was desired to maximize information gathered in this study, minimal amounts of
165 understory vegetation deemed obstructive were manually removed.

166 Scanning was conducted at a speed of 122,000 pulses per second, resulting in approximately
167 seven minutes duration per scan. With transport and setup time between scan positions taking
168 an average of 2-3 minutes, scanning each tree from all three angles took on average 25-30
169 minutes.

170 The scan data were collected in a local coordinate system using the scanner location as the
171 origin. Registration was done automatically using SCENE v4.8 software (FARO 2014) based
172 on printed checkerboard targets placed within each scan image. Quality of registration was
173 reported as average discrepancy in distance between a given pair of reference objects or
174 tension (ranging between 1mm to 8mm). For each registered scan, TLS returns belonging to
175 an individual sample tree were selected visually from the 3D representation of the
176 surrounding forest. This process was done by displaying the registered scan and using the
177 visible scan positions to deduce which was the sample tree (Figures 2 and 3). Prior to
178 scanning, boles of the sample trees were wrapped with very thin striped plastic flagging,
179 intentionally placed well above DBH, that proved visible as a final confirmation the correct

180 tree was to be selected from the registered point cloud. These selected points were then
181 exported for later use in extracting DBH, HT and HTCB using Matlab 2013b (The
182 MathWorks, <http://www.mathworks.com>, USA).

183 *2.4. Tree Parameter Extraction from Selected Scan Data*

184 TLS based height measurements were normalized to the surface elevation by means of a
185 digital terrain model (DTM) derived from the TLS data. Ground and non-ground returns were
186 separated using a grid based approach to select the lowest return within a $0.3048\text{m} \times 0.3048\text{m}$
187 sampling grid placed over the plot area.

188 *2.4.1. Tree Detection*

189 With the ground model complete the next step before obtaining tree parameters was to
190 estimate the center of the sample tree at approximately 1.37m (diameter at breast height)
191 above the ground. This estimated location served as a control point from which all
192 measurement algorithms originated from. Similar to Mass et al. (2008) a thin 5-10cm
193 horizontal slice was selected from the point cloud for stem detection (Figure 2). This
194 horizontal slice often included many points representing branches and foliage at that height.
195 To expedite the estimation process only a subset of the points in the slice was used (Figure 3).
196 A nonlinear least squares circle-fitting procedure, similar to that described by Henning and
197 Radtke (2006), was used for estimating the diameter and XY center of each tree. The means
198 of the XY coordinates of all subset points were used as initial estimates, or starting values, for
199 the nonlinear procedure, provided there were no large outliers in the point cloud (Maisonobe
200 2007). Restriction of the subset to the XY range of the main bole additionally addresses any
201 outliers associated with branches or foliage. The starting value for the diameter of the sample
202 tree consisted of using the following equation to solve for a diameter for each of the subset

203 points, and then using the mean of all calculated diameters, produced using the following
 204 equation (Henning and Radtke 2006):

$$\hat{d}_i = 2 \times \sqrt{(\hat{x}_c - x_i)^2 + (\hat{y}_c - y_i)^2} \quad (1)$$

205

206 where \hat{d}_i is estimated diameter for the i^{th} subset point, the (\hat{x}_c, \hat{y}_c) pair are the means of the
 207 (x,y) coordinates for all subset points and the (x_i, y_i) pair are the (x,y) coordinates of the i^{th}
 208 subset point. With these three starting values and the following equation as the objective
 209 function, the three unknowns were solved for by minimizing the sum of squares for all subset
 210 points:

$$F_i = 2 \times \sqrt{(\hat{x}_c - x_i)^2 + (\hat{y}_c - y_i)^2} - \hat{d}_i \quad (2)$$

211 where F_i is the value of the objective function for the i^{th} subset point and \hat{x}_c, \hat{y}_c and \hat{d}_i are the
 212 three unknowns. With the spatial location of the center of the tree and its diameter
 213 approximated, subsequent measurements stemmed from this information.

214 2.4.2. Uphill Side of the Tree and Total Height

215 To conform to forest inventory practices, all heights up the bole to the tip of the sample trees
 216 were measured relative to the ground adjacent to the tree with the highest elevation (Avery
 217 and Burkhart 2002, p.144). Thus, it was necessary to identify the uphill-side of the tree and
 218 determine the elevation of that side relative to the rest of the point cloud. Using the
 219 approximated center and diameter, all DTM cells determined to be spatially adjacent to the
 220 base of the sample tree were selected. The selected DTM cell with the highest elevation value
 221 was determined to be the uphill-side of the tree. The corresponding elevation value of that

222 cell, heretofore referred to as the reference z-value, was used as the minimum reference
223 height for extracting DBH, HT and HTCB.

224 A statistical quality control was implemented in order to ensure the reference z-value was not
225 a far outlier representing anomalies such as nearby rocks or protruding tree roots. Whereby, if
226 the coefficient of variation (CV) of the elevation values of all the selected adjacent DTM
227 cells was above a defined percentage, the adjacent cell with the next highest elevation value
228 was chosen from the eight cells that bordered the stem center grid cell. For the purpose of this
229 study a subjectively chosen CV of 60% was used to remove outlier values. HT was then
230 simply calculated as the difference between the highest point in the point cloud and the
231 reference z-value. Stray points above the tip of the tree were not observed to be a problem
232 due to the prior filtering using the SCENE software.

233 2.4.3. Diameter at Breast Height

234 While the approximated diameter from the previously-described detection slice at breast
235 height could potentially serve as an estimate of DBH, the height at which the slice was taken
236 was 1.37m above the minimum elevation of the entire point cloud, rather than the uphill side
237 of the tree. On steeply sloping terrain, differences in relative bole heights could be
238 substantial. To avoid this issue, an improved DBH was extracted 1.37m above the reference
239 z-value using the previously described procedure of subsetting followed by the non-linear
240 least squares circle-fitting. However, an additional precision constraint was added to
241 maximize the reliability of the DBH measurement. If the root mean square error (RMSE) of
242 the non-linear least squares procedure was above a defined threshold of 5mm, a recursive
243 “noise reduction” method, similar to Henning and Radtke (2006), was invoked. The main
244 purpose of this procedure was to reduce TLS observations originating from nearby branches
245 or understorey. Henning and Radtke (2006) showed the removal of these outliers greatly

246 improved our estimates of DBH. The filtering process involved continually removing the
247 points whose coordinates produced estimated diameters that were the maximum absolute
248 distance from the mean of all estimated diameters until the standard deviation of the
249 estimated diameters was below the same defined threshold. It was observed that using 5mm
250 for this threshold was sufficient for minimizing the measurement error, while also removing
251 stray points around, and not belonging to, the main bole.

252 2.4.4. Height to Crown Base

253 Estimation of HTCB was based on analysis of point intensity and percentiles of return height.
254 Point intensity is a measure of the returned energy of an emitted pulse. While intensity values
255 cannot directly be used as a surrogate to optical measures of surface reflectance (as these
256 uncalibrated measurements depend on environmental conditions, scanner properties and
257 location) LiDAR based intensity measures have been successfully used to distinguish
258 between green foliage and non-photosynthetically active tree elements (Popescu et al. 2007,
259 Pesonen et al. 2008 and Kim et al. 2009), due to the large differences in NIR reflectance of
260 these vegetation components. As a result, we used intensity measures to classify between
261 foliage and woody surfaces for the purpose of estimating HTCB. By plotting intensity versus
262 height, an empirical threshold could be determined, below which the intensity values for
263 points returned from foliage would theoretically occur (Figure 4). The subset of points below
264 this threshold served as a representation of the live crown profiles. We then used different
265 percentile heights within this subset of points for different age classes of the trees.
266 Specifically, the 5th, 10th and 25th percentile height of this subset were used to measure
267 HTCB for the 20-40yr, 40-80yr and >80yr age classes of sample trees selected, respectively.

268 2.5. Modeling biomass

269 The biomass equations used for this study predict the proportion of AGB for the bole, bark,
 270 branch and foliage component (Poudel and Temesgen 2016 and Poudel et al. 2015). These
 271 proportions can then be multiplied by an estimate of total tree AGB to obtain the AGB of
 272 each tree component. Both the component equations and the total tree biomass equation were
 273 fit in separate systems of equations using the seemingly unrelated regression method (SUR)
 274 in SAS statistical software (SAS Institute Inc., v9.4). The four CRM component equations
 275 and the total tree equation used in our study are of the form (Poudel and Temesgen 2016 and
 276 Poudel et al. 2015):

$$p\widehat{Bole}_i = \exp[\widehat{\beta}_0 + \widehat{\beta}_1 \times \ln(DBH_i) + \widehat{\beta}_2 \times \ln(HT_i) + \frac{\widehat{\sigma}_1^2}{2}] \quad (3)$$

$$p\widehat{Bark}_i = \exp[\widehat{\beta}_3 + \widehat{\beta}_4 \times \ln(DBH_i) + \frac{\widehat{\beta}_5}{\ln(HTCB_i)} + \frac{\widehat{\sigma}_2^2}{2}] \quad (4)$$

$$p\widehat{Branch}_i = \exp[\widehat{\beta}_6 + \widehat{\beta}_7 \times \ln(DBH_i) + \widehat{\beta}_8 \times \ln(HTCB_i) + \frac{\widehat{\sigma}_3^2}{2}] \quad (5)$$

$$p\widehat{Foliage}_i = \exp[\widehat{\beta}_9 + \widehat{\beta}_{10} \times \ln(DBH_i) - \widehat{\beta}_{11} \times \ln(HTCB_i) + \frac{\widehat{\sigma}_4^2}{2}] \quad (6)$$

$$\widehat{Total\ Tree}_i = \exp[\widehat{\beta}_{12} + \frac{\widehat{\beta}_{13}}{DBH_i} + \frac{\widehat{\sigma}_5^2}{2}] \quad (7)$$

277 where \bar{p}_{Bole} , \bar{p}_{Bark} , \bar{p}_{Branch} , and \bar{p}_{Foliage} are the estimated proportions of component
 278 AGB for bole wood, bark, branches and foliage, respectively, $\exp(\cdot)$ is the exponential
 279 function, $\ln(\cdot)$ is the natural logarithm function and the $\hat{\beta}_i$ are the estimated parameters from
 280 the SUR procedure. The $\frac{\sigma^2}{2}$ is the correction factor for the resulting bias when back-
 281 transforming model predictions from the logarithmic to the initial scale of interest, where $\hat{\sigma}^2$
 282 is the estimated mean squared error, or residual variance (Baskerville, 1972, McRoberts and
 283 Westfall 2014).

284 2.6. Measurement Error Variability

285 For DBH, HT and HTCB, the differences between the measured values and the downed-tree
 286 measurements were calculated for both the TLS and traditional forest inventory instruments.
 287 In this study, the downed-tree measurements were considered to be the known “true” values
 288 due to the ease with which measurements could be taken as accurately as possible. The
 289 summary data for these differences were subsequently calculated for each input variable for
 290 the models (Table 2).

291 It is known that standard deviation of the measurement error is zero when HT is zero. Hence,
 292 to stay consistent with the methodologies of Berger et al. (2014) and Shettles et al. (2015) a
 293 simple linear regression model through the origin was constructed to predict the standard
 294 deviation of the measurement errors. In order to conduct regressions of standard deviation of
 295 measurement errors on input variables, HT values were sorted in ascending order and
 296 grouped into groups of size 3, with the last group including the remainder of the HT values.
 297 With an aim to maximize the number of possible groups, the group size of 3 was
 298 symptomatic of a sample size of 25 trees. For every g^{th} group, the means of the HT values

299 and SD_{ME_g} were estimated, where $SD_{ME_g} = \frac{1}{n-1} \sqrt{\sum_{i=1}^n (ME_{HT,g} - \overline{ME_{HT,g}})^2}$ is the

300 standard deviation of the measurement errors for HT and $ME_{HT,g} = HT_D - HT_S$ are the HT
 301 measurement errors, where HT_D is the downed-tree height measurement, n is the group size
 302 and HT_S is the standing-tree height measurement. The following model form was fit to the
 303 grouped data for HT using the method of ordinary least squares:

(8)

$$\widehat{SD}_{ME,HT} = \hat{\beta}_{14} \times HT$$

304 where $\widehat{SD}_{ME,HT}$ is the estimated standard deviation of the measurement errors for HT and $\hat{\beta}_{14}$
 305 is the model parameter estimate.

306 2.7. Integrating Simulated Measurement Errors into Model Uncertainty

307 Using the standard deviations from Table 2 and equation 8, Monte Carlo simulations (>5000
 308 iterations), were used to approximate model uncertainties reflective of the additional
 309 uncertainty due to measurement error (Berger et al. 2014). Input variable contamination was
 310 implemented as a two part process: First, for the kth component model, a multiplicative factor
 311 $\sim N(1, SD_{HT}^2)$ was randomly generated and multiplied together with the input variables, where
 312 SD_{HT} is the standard deviation of the height measurement errors; and second, an additive
 313 factor $\sim N(0, \widehat{SD}_{ME,HT}^2)$ was randomly generated and added to the input variables, where
 314 $\widehat{SD}_{ME,HT}$ is the predicted standard deviation from equation 8 (Berger et al. 2014).

315 The impact of the additional uncertainty was assessed by calculating the mean prediction and
 316 RMSE and the relative RMSE (RRMSE) over all iterations with the following formulas:

$$mcan = \frac{1}{n} \sum_{i=1}^n \hat{Y}_i \quad (9)$$

$$\text{RMSE} = \sqrt{\frac{1}{n} \sum_{i=1}^n (Y_i - \hat{Y}_i)^2} \quad (10)$$

317 where Y_i is the observed value and \hat{Y}_i is the fit for the i^{th} tree. RRMSE is calculated by simply
 318 dividing RMSE by the mean.

319 To convert the predicted proportions and RMSEs to tree-level units (oven-dry kg), the
 320 predicted proportions were multiplied by the fitted value for total tree biomass to obtain tree-
 321 level fitted values of component AGB (Eq. 7), and multiplied with the absolute RMSEs
 322 produced as the square root of the sum of the squared relative RMSEs:

$$\mathcal{E}AGB_{\text{Comp}} = \overline{AGB}_{\text{Comp}} \times \sqrt{\left(\frac{\mathcal{E}AGB_{\text{Ratio}}}{\overline{AGB}_{\text{Ratio}}}\right)^2 + \left(\frac{\mathcal{E}AGB_{\text{TT}}}{\overline{AGB}_{\text{TT}}}\right)^2} \quad (11)$$

323 Where $\mathcal{E}AGB_{\text{Comp}}$ is the combined RMSE in tree-level units, $\mathcal{E}AGB_{\text{Ratio}}$ is the RMSE for the
 324 CRM component ratios and $\mathcal{E}AGB_{\text{TT}}$ is the RMSE for Total Tree AGB (equation 7).

325 2.8. Integrating Model Error into Sampling Uncertainty

326 In order to integrate the model errors into the sampling uncertainty, the magnitude of the
 327 model errors integrated needed to be contingent upon the magnitude of the model predictions.
 328 Using the previously described grouping approach with respect to the model errors, a simple
 329 linear regression model (also forced through the origin) was constructed to predict the
 330 magnitude of the model errors. Following the notation and general methodology of
 331 McRoberts and Westfall (2014): (1) for the k^{th} component model, a joined list of ϵ_i , Y_i and \hat{Y}_i
 332 was created and sorted in ascending order with respect to \hat{Y}_i , where $\epsilon_i = \hat{Y}_i - Y_i$; (2) the sorted
 333 triads of observations were grouped into groups of size 3, with the last group including the
 334 remainder of the observations; (3) for every g^{th} group, the mean observation $\bar{Y}_g = \frac{1}{n_g} \sum_{i=1}^{n_g} Y_i$,

335 the mean fitted value $\bar{Y}_g = \frac{1}{n_g} \sum_{i=1}^{n_g} \hat{Y}_i$ and the mean square error $\sigma_g^2 = \frac{1}{n_g-1} \sum_{i=1}^{n_g} \varepsilon_i^2$ were
 336 calculated, where n_g is the number of trees in the g^{th} group; (4) the following model form was
 337 fit to the grouped data for each component model using the method of ordinary least squares

$$\hat{\sigma}_i = \hat{\beta}_{15} * \hat{Y}_i \tag{12}$$

338 Where $\hat{\sigma}_i$ is the predicted model error for the i^{th} tree, $\hat{\beta}_{15}$ is the model parameter estimate and
 339 \hat{Y}_i is the model fitted value for the i^{th} tree. It should be noted that with measurement error
 340 integrated into the model errors, the value of $\hat{\beta}_{15}$ is expected to increase, reflecting this
 341 additionally accounted for source of uncertainty.

342 A bootstrapping technique, in conjunction with equation 12, was used to simulate the effects
 343 of model errors on the uncertainty of per unit area estimates of component AGB for all
 344 models. A similar Monte Carlo simulation sequence and notation described by McRoberts
 345 and Westfall (2014) was used for each component model.

346 First, the data set containing the “true” values of the 25 sample trees was randomly sampled
 347 with replacement to produce a bootstrapped-sample of size 25. Similar to the previously
 348 described method of simulating measurement errors, contaminated model predictions for all
 349 25 bootstrapped-sampled trees were produced by adding a randomly generated residual, ε_i
 350 $\sim N(0, \hat{\sigma}_i^2)$, to the prediction for the i^{th} pseudo-sampled tree produced using the k^{th} component
 351 model, where $\hat{\sigma}_i$ is estimated using equation 12. Using the contaminated predictions and the
 352 pseudo sample data, a new model, of the same form as the k^{th} component model, was refit.
 353 For equations 3, 4, 5, 6 and 7, due to their original model form, the contaminated predictions

354 and the pseudo sample data required transformation to the \log_{10} - \log_{10} and ln-ln scale,
 355 respectively, prior to refitting.

356 Second, the refit equations were applied to the ground plot data set. For the i^{th} tree in the j^{th}
 357 plot, predictions of tree-level component AGB were produced by adding the model
 358 predictions to a randomly generated constrained residual, $\lambda \epsilon_i$ where ϵ_i is the randomly
 359 generated residual $\sim N(0, \hat{\sigma}_i^2)$, and λ is a multiplicative constraining factor that yields model
 360 efficiency values of 0.95. Model efficiency, calculated as

$$Q^2 = 1 - \left(\frac{\sum_{i=1}^{n_{pl}} \epsilon_i^2}{\sum_{i=1}^{n_{pl}} (Y_i - \bar{Y})^2} \right) \quad (13)$$

361 where n_{pl} is the number of trees in the ground plot data set, is a goodness-of-fit statistics
 362 similar to the coefficient of determination from the ordinary least squares procedure, where
 363 the higher the value the better the fit of the model to a given data set (Vanclay and
 364 Skovsgaard 1997, McRoberts and Westfall 2014). This multiplicative factor constraint was
 365 implemented in order to have a standardized quality of fit of the model to the ground plot
 366 data for purposes of comparing the standard errors of the mean for all component models.
 367 Due to recent published findings illustrating the minimal effect correlation among trees
 368 within plots has on the standard error of the estimates, correlation among residuals was not
 369 integrated into the analysis of this study (Berger et al. 2014, Breidenbach et al. 2014,
 370 McRoberts et al. 2014).

371 Third, to obtain the estimated per hectare values of component AGB on the j^{th} cluster plot, the
 372 summation of all subplot-level per unit area component AGB predictions on the l^{th} subplot
 373 were calculated as

$$Y_j = \sum_{l=1}^4 Y_l \quad (14)$$

374 with

$$Y_l = \frac{\sum_{i=1}^{n_l} Y_{i,l}}{\text{Subplot Area}_{\text{hectares}}} \quad (15)$$

375 Where n_l is the number of trees observed in the l^{th} subplot and $Y_{i,l}$ is the i^{th} tree on the l^{th}
 376 subplot. Fourth, for each simulation cycle the mean and variance of the mean across all
 377 cluster plots were calculated as

$$\bar{Y} = \frac{1}{n_{cl}} \sum_{j=1}^{n_{cl}} Y_j \quad (16)$$

$$\widehat{\text{Var}}(\bar{Y}) = \frac{1}{n_{cl}(n_{cl}-1)} \sum_{j=1}^{n_{cl}} (Y_j - \bar{Y})^2 \quad (17)$$

378

379 Where n_{cl} is the number of cluster plots (25 in this study). Finally, the mean prediction and
 380 mean within-simulation variance over 5000 simulation cycles were calculated as

$$\hat{\mu}_{sim} = \frac{1}{5000} \sum_1^{5000} \bar{Y} \quad (18)$$

$$\widehat{\text{Var}}_{sim} = \frac{1}{5000} \sum_1^{5000} \widehat{\text{Var}}(\bar{Y}) \quad (19)$$

381 Comparisons of the mean predictions as well as final propagated error were compared for all
 382 component models for both approaches. Metrics used for comparison included RMSE,
 383 RRMSE, standard error of the mean (SE) from equation 19 and relative SE (RSE).

384 **3. Results and Discussion**

385 *3.1. Measurement Errors*

386 Table 2 shows the measurement error summary statistics for input variable measurements
387 using the TLS and the STM. The circle-fitting procedure for measuring DBH resulted in 6 of
388 the 25 trees showing agreement with the downed tree measurements, and 9 being within 3cm.
389 These results are comparably better than previous studies assessing the quality of TLS-
390 derived diameter measurements. Simonse et al. (2003) used a Hough-transformation to obtain
391 DBH for 23 trees, reporting minimum, maximum, mean and standard deviation of
392 measurement error values as -5.8cm, 5.6cm, 1.7cm and 2.8cm, respectively. Hopkinson et al.
393 (2004) reported an average difference of 10cm for plot-level comparison of DBH between
394 TLS and manual measurement techniques. Thies et al. (2004) used a stem reconstruction
395 method involving the fitting of a series of cylinders up the main stem of two scanned
396 deciduous trees of different species. DBH was calculated as the diameter of the
397 corresponding cylinder at breast height. Deviations in TLS-derived DBH measurements from
398 standing tree measurements were -1.3cm and 0.6cm for European beech and wild cherry,
399 respectively. Henning and Radtke (2006) reported errors of less than 1cm (0.3in) using a
400 similar circle-fitting procedure as the one described here when comparing TLS diameters to
401 known values from felled trees. In a separate study attempting to model 3D plot-level forest
402 structure, Henning and Radtke (2006) reported an average DBH difference of 4.8cm when
403 comparing TLS measurements to standing tree measurements. Most likely, the quality of our
404 TLS-derived DBH results compared to other studies is largely attributable to our multi-scan
405 approach, which has been shown to reduce the variability TLS-derived DBH measurements
406 by drastically increasing the cover of point clouds (Pueschel et al. 2013). Using a multi-scan
407 dataset and quantitative structure models to obtain inferred ABG through estimated total
408 height and DBH, Calders et al, 2014 reported a concordance correlation coefficient of 0.98. A
409 RMSE threshold below 5mm often resulted in underestimations of DBH. Presumably, this

410 was due to points on the outside of the fissures of the bark being the points removed first
411 during this point removal process. Because the true values of DBH were measured on the
412 outside of these fissures, stricter thresholds were not used. Hence, if this procedure is to be
413 used for older trees of a species with deeply fissured bark characteristics, this process may
414 require allowing for higher RMSE thresholds. Average RMSE observed for the fitting of all
415 25 DBHs was 3.99 mm. This process holds promise for obtaining upper stem diameters
416 outside bark for purposes of taper determination, form factor calculation and possible
417 merchantable height identification as well. While more robust methods exist for stem
418 detection and outlier determination that do not rely on a circularity tolerance, these results are
419 still relevant to assessing uncertainty in AGB estimates obtained through TLS, a topic studied
420 little up to this point.

421 HT measurement error results for TLS showed lower average bias than the STM HT
422 measurements at -0.1m and -1.0m for TLS and STM, respectively. Encouragingly, the
423 standard deviation of these measurement errors for HT was also lower for TLS, at 0.3m and
424 0.7m for TLS and STM, respectively. These estimates are lower than those reported by
425 Hopkinson et al. (2004), who reported an average difference of 1.5m for plot-level HT
426 comparisons between TLS and manual measurement techniques. Their reported difference in
427 standard deviations of HT measurements was lower at 0.2m. Chasmer et al (2006) reported
428 an average underestimation of HT of 1.2m for 15 trees within a closed-canopy stand of red
429 pine (*Pinus resinosa*) scanned from five different locations. The comparative improvement
430 upon these studies suggests this method of identifying a reference z-value from which to
431 subtract from the maximum z-value is superior to other methods. However, with stand
432 density and tree size being limiting factors in the accuracy of TLS-derived-HT
433 measurements, the quality of the results we present here for HT could also likely be a result
434 of several of the sample trees being from stands with lower stand densities, and lodgepole

435 pine being a relatively shorter tree species. The capability of the FARO Focus^{3D} 120 to scan
436 at the point density chosen for this study also likely furthered this improvement.

437 In contrast to HT, HTC_B results for the TLS exhibited a larger mean and standard deviation
438 of the measurement errors compared to the STM. However, this variable has typically been a
439 point of imprecision for TLS extraction procedures. Thies et al. (2004) reported differences in
440 HTC_B values of -0.12m and -0.11m for the two aforementioned sample trees. With the
441 sample trees being relatively large, forked and deciduous, HTC_B was measured as the height
442 to the first fork. Jung et al. (2011) compared HTC_B measurements from coincident ALS and
443 TLS data, where the TLS measurements were considered to be the actual values. ALS HTC_B
444 values were obtained using k-means clustering technique which groups the point cloud into a
445 user-defined number of classes based upon differences in the spatial distribution of points
446 within the point cloud. The authors chose three classifications to represent ground cover,
447 understory vegetation and canopy cover. ALS HTC_B was determined from the lowest point
448 in the canopy cover classification. Because differences in the point density distribution were
449 deemed too small with the TLS data, k-means clustering was not used, replaced by manual
450 identification of the lowest crown return via a monitor display. The difference in mean HTC_B
451 values was reported as 0.2m.

452 Using the height of the lowest point in this subset as a measure of HTC_B resulted in
453 consistent underestimation, similar to the results observed by Chasmer et al. (2006). This was
454 likely due to: (1) the presence of dead branches interspersed within the lower portion of the
455 live crown, as is common for lodgepole pine; and (2) the definition of HTC_B used in this
456 study being the height to the lowest live limb rather than the height to the lower margin of the
457 main live crown. Thus, HTC_B was then estimated as the 5th, 10th and 25th percentile height of
458 this subset for the 20-40yr, 40-80yr and >80yr age classes of sample trees selected,
459 respectively. Selection of this threshold was based upon: (1) empirical observation; and (2)

460 the knowledge that younger lodgepole pine trees typically have lower HTCB values and
461 fewer dead branches. Due to this method yielding the lowest average measurement error, the
462 measurements resulting from this approach were ultimately selected for use in the subsequent
463 error propagation analysis.

464 Our results show that applying the TLS inventory parameter extraction techniques to
465 inventory applications could be a useful approach to complement conventional data
466 acquisition techniques; however, further validation will be needed for broader scale
467 applications across different forest types/larger areas. We acknowledge that sampling
468 conditions and sample size of 25 destructively measured trees is not sufficiently
469 representative to extrapolate our findings across larger areas or different vegetation types.
470 Our approach should therefore be understood as a first demonstration of error and error
471 propagation obtainable from terrestrial laser scanning using ground data in PNW forests.

472 Future improvements would further bolster the applicability of TLS to larger operations.
473 First, rather than the manual graphical method for tree detection employed here, more
474 sophisticated automatic tree detection procedures that omit non-bole points from branches
475 and foliage, would be necessary. Secondly, for the subsetted percentile approach for HTCB,
476 the tree size to percentile relationship may need to be more generalized by diameter classes,
477 or calibrated to the specific operation.

478 *3.2. Model Predictions and Uncertainty*

479 When the measurement error was integrated into the CRM equations, mean predictions of
480 AGB for all components were similar between instruments (Table 4). The TLS RMSE values
481 for the CRM ratios were lower for all components compared to the STM RMSE values
482 (Table 5). However, the RMSE values were larger for the TLS, primarily due to the Total
483 Tree SUR equation having a 147% larger RMSE. This can be attributed to the assumption

484 that the STM measurement of DBH, the only input variable for the SUR equation, was
485 measured without error, hence the STM simulation procedure did not involve the
486 contamination of DBH values. Had the measurement error in DBH been assessed, it is
487 feasible to reason that the uncertainty value for the STM SUR equation would have been
488 greater than the 70.59kg value reported in this study. Also worthy of noting, the magnitude of
489 the difference between STM and TLS may have been less dramatic when the total tree
490 equation would have been fit in a common system of equations. Nevertheless, the tree-level
491 uncertainty in predicted component AGB associated with using the TLS for extracting input
492 variables for the CRM equations is likely greater than estimates using a spencer tape by
493 trained individuals.

494 *3.3. Per Unit Area Estimates and Uncertainty*

495 With model errors incorporated into the simulations for per unit area estimation, the
496 uncertainty increased markedly (Tables 6 & 7). This notable increase is further illustrated in
497 Table 9, which shows that the RSE values for all components increased two to three-fold.
498 Encouraging, however, was the notable difference in per unit area precision between
499 instruments when measurement error was integrated. (Table 9). The relative proportions of
500 SE due to measurement, model and sampling error using SRM were 11%, 66% and 23%,
501 respectively. The relative proportions of SE due to measurement, model and sampling error
502 using the TLS were 5%, 70% and 25%, respectively. Improvements in the measurement error
503 from TLS were largely the result of increased accuracy of tree height as well as height to
504 crown base. We acknowledge that the manual nature of the vegetation removal and the
505 extraction of individual trees from co-registered scans possibly resulted in optimistic values
506 of error contribution from the TLS. Nonetheless, our findings suggest using the TLS can

507 result in a lower propagated error, primarily due to a smaller contribution to the total
508 uncertainty from measurement error.

509 **4. Conclusion**

510 With broad-scale inventories, such as FIA and others likely to face an increased demand for
511 defensible AGB uncertainty estimates, accounting for and addressing all primary sources of
512 error becomes paramount. Taking the Monte Carlo approach shown here, measurement and
513 model error have been successfully integrated and accounted for. With only 25 subjectively
514 selected trees for use in comparison, the inference made here is an approximation. However,
515 not only were the general contributions for all three sources of error illustrated, the addressal
516 of measurement error was made by showing that the use of the TLS indeed can improve
517 precision of per unit area estimation of lodgepole pine AGB using the component equations
518 presented here.

519 Future research into this matter could also be best directed at similarly assessing the
520 propagated error from using the TLS with other AGB models, as well as models for other
521 parameters of interests, both point-in-time and growth-related. The TLS data analysis
522 techniques shown here hold value in reducing uncertainty attributed to measurement error,
523 which has been shown to contribute a potentially serious amount to the total per unit area
524 uncertainty AGB estimates. Investigations into using the same multi-scan approach for plot-
525 level analysis would add credence to the work done here, as that is likely the more applicable
526 inventory scenario forest managers would be utilizing the TLS, rather than for single trees, as
527 was done in this study. Extraction of additional tree-level input variables, such as upper stem
528 diameters, merchantable top height and crown width would provide additional information
529 about how the performance of the TLS in extracting these variables propagates up to per unit
530 estimates of AGB. All of these future research efforts are likely to increase the defensibility

531 of reported precision estimates for AGB derived using individual-tree equations, while also
532 helping determine under which scanning scenarios, and for which input variables, does the
533 use of the TLS translate into quantifiable gains in precision for broad-scale estimates of
534 AGB.

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543

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658 in dry kilograms of biomass per hectare.

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660

661

662 Table 1: TLS technical data

Specification	Focus ^{3D} 120
Range Finder	Phase shift
Field of view (horizontal x vertical)	360° x 305°
Measurement range	0.6m – 120m
Distance accuracy	± 2mm at 25m
Sampling Rate	Up to 976k/sec
Beam radius at discharge	3.0mm
Beam divergence	0.19mrad (0.011°)
Weight	5.0kg

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665 Table 2: Summary statistics of the measurements errors for STM and TLS

Standing Tree Measurements (STM)					
	n	Min.	Mean	Max.	SD
HT (m)	25	-2.56	-0.98	0.12	0.67
HTCB					
(m)	25	-1.04	-0.06	1.37	0.52
Terrestrial LiDAR (TLS)					
	n	Min.	Mean	Max.	SD
DBH					
(cm)	25	-1.27	-0.25	1.52	0.51
HT (m)	25	-0.79	-0.06	0.55	0.27
HTCB					
(m)	25	-3.29	0.49	3.90	1.68

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669 Table 3: Model predictions and RMSE values for CRM ratios and CRM tree-level estimates without measurement error. Tree-levels units are in kilograms of
 670 dry biomass.

Model Means-Without Measurement Error			Model RMSEs-Without Measurement Error		
Total Tree (SUR)	287.11		Total Tree (SUR)	70.59	
	CRM	CRM Tree-			
Component	Ratios	Level	Component	CRM Ratios	CRM Tree-Level
Bole	0.672	193.07	Bole	0.067	51.24
Bark	0.055	15.66	Bark	0.034	10.62
Branch	0.195	56.04	Branch	0.055	21.02
Foliage	0.082	23.40	Foliage	0.022	8.63

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673 Table 4: Model predictions for CRM ratios and CRM tree-level estimates with measurement error, for STM and TLS. Tree-levels units are in kilograms of dry
 674 biomass.

Model Means-With Measurement Error				
Total Tree (SUR)-STM	287.11			
Total Tree (SUR)-TLS	296.57			
	STM	TLS	STM	TLS
	CRM	CRM	CRM Tree-	CRM Tree-
Component	Ratios	Ratios	Level	Level
Bole	0.599	0.620	172.03	177.97
Bark	0.061	0.061	17.45	17.47
Branch	0.209	0.208	60.11	59.80
Foliage	0.091	0.091	26.19	26.21

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679 Table 5: Model RMSE values for CRM ratios and CRM tree-level estimates with measurement error, for STM and TLS. Tree-levels units are in kilograms of
 680 dry biomass.

Model RMSEs-With Measurement Error				
Total Tree (SUR)-STM		70.59		
Total Tree (SUR)-TLS		174.70		
	STM	TLS	STM	TLS
Component	CRM Ratios	CRM Ratios	CRM Tree-Level	CRM Tree-Level
Bole	0.297	0.067	95.30	123.09
Bark	0.047	0.034	14.04	15.16
Branch	0.074	0.055	25.81	43.16
Foliage	0.037	0.022	12.46	19.19

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682

683 Table 6: Per hectare estimates and SE values for CRM equations, without accounting for measurement or model error. Units are in kilograms of dry biomass
684 per hectare.

Sampling Error Only		
Component	Mean	SE
Bole	23,270.38	4,945.90
Bark	1,842.11	357.83
Branch	6,414.97	1,234.54
Foliage	2,544.31	463.86

685 Table 7: Per hectare estimates and SE values for CRM equations accounting for model error. Units are
 686 in dry kilograms of biomass per hectare.

Sampling Error (With Model Error)		
Component	Mean	SE
Bole	37,062.80	17,659.36
Bark	4,947.56	3,317.42
Branch	9,275.29	4,043.61
Foliage	3,425.94	1,234.22

687

688 Table 8: Per hectare estimates and SE values for CRM equations accounting for model error. Units are
 689 in dry kilograms of biomass per hectare.

Sampling Error (With Model and Measurement Error)					
Mean			SE		
Component	STM	TLS	Component	STM	TLS
Bole	37,819.12	37,442.94	Bole	19,201.47	18,583.77
Bark	3,984.19	3,740.49	Bark	4,502.45	3,429.02
Branch	9,475.49	9,475.52	Branch	4,266.80	4,363.27
Foliage	3,512.56	3,547.24	Foliage	1,316.67	1,355.31

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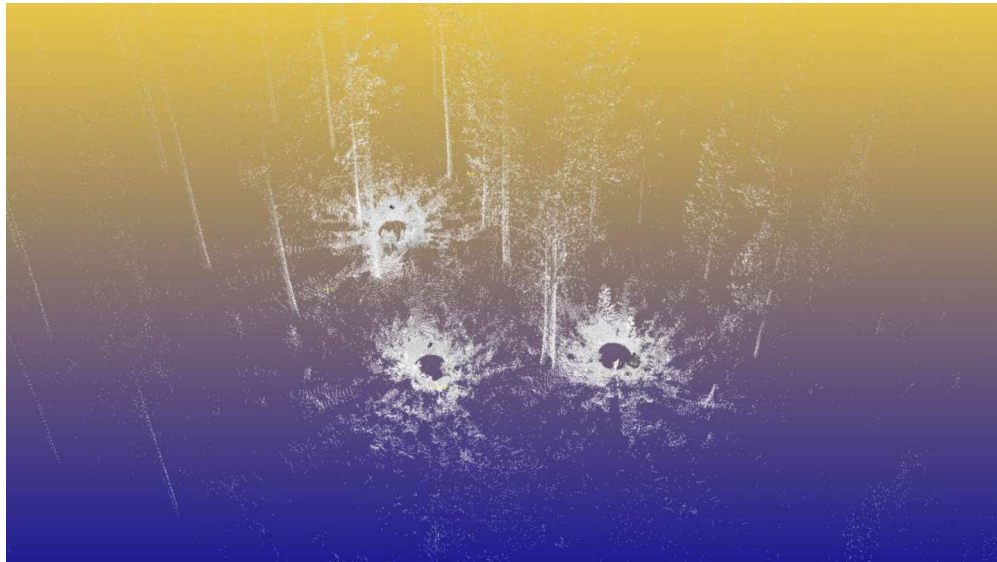
696 Table 9: RSE values for CRM equations accounting for model and measurement error.

Sampling Error (RSEs)				
Component	Sampling Only	Model Errors	Model and Measurement Errors	
			STM	TLS
Bole	21.3%	47.6%	50.8%	49.6%
Bark	19.4%	67.1%	113.0%	91.7%
Branch	19.2%	43.6%	45.0%	46.0%
Foliage	18.2%	36.0%	37.5%	38.2%

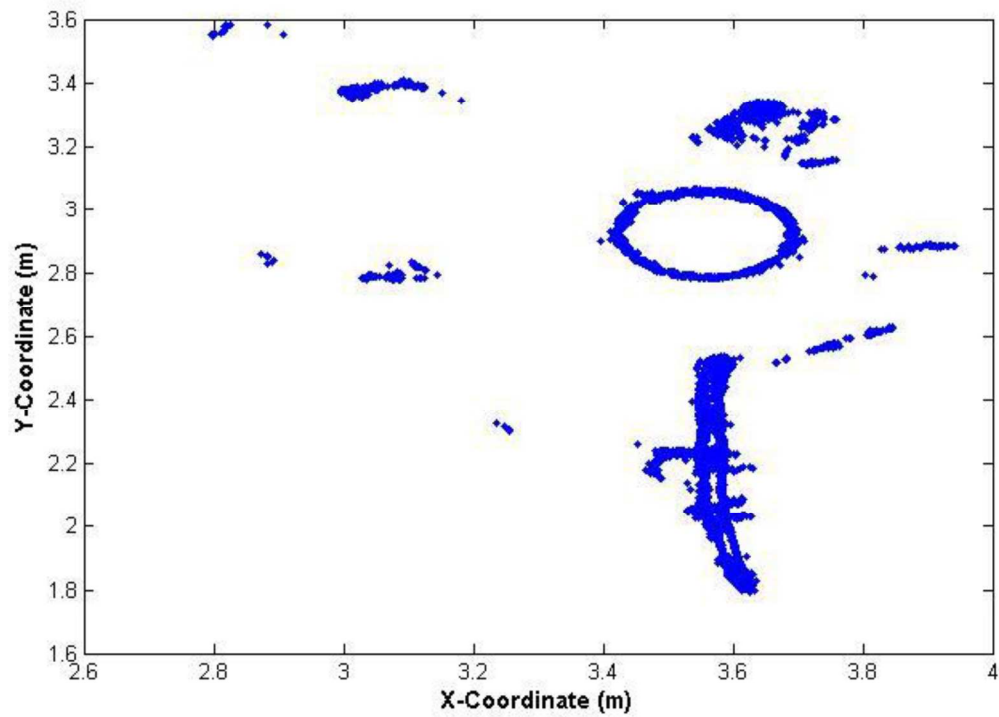
697

698

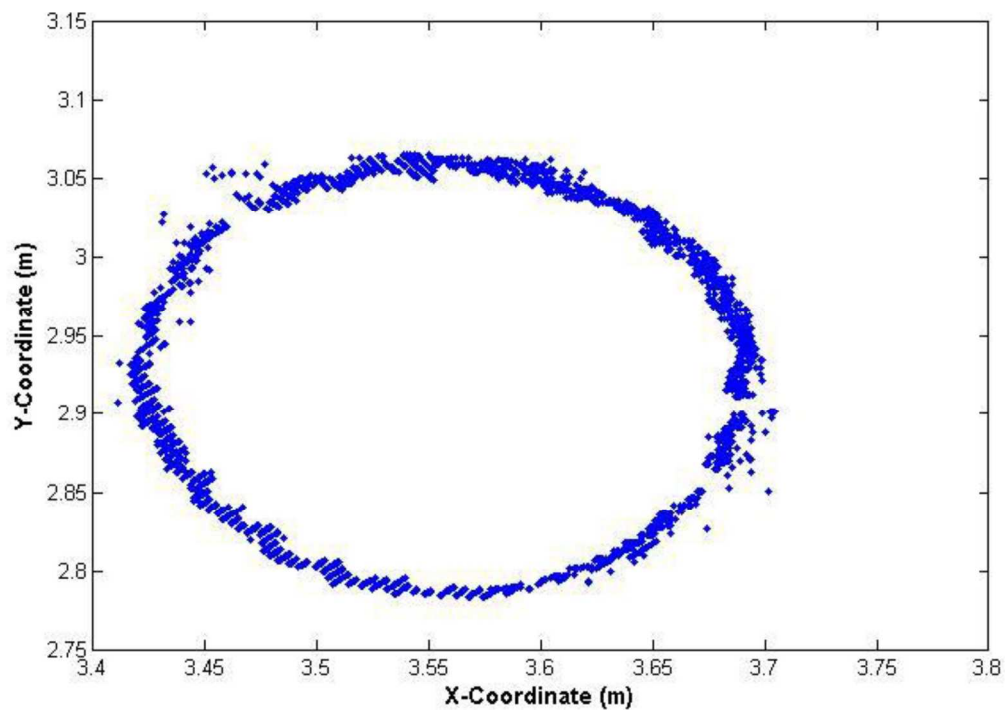
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703 Unrestricted subset included branches and foliage located within the height range of the slice.
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707 subset threshold is shown in red.
- 708
- 709



Filtered overhead 3D view of registered point cloud. Black circles denote scan locations around the sample tree, located right center.
839x469mm (72 x 72 DPI)

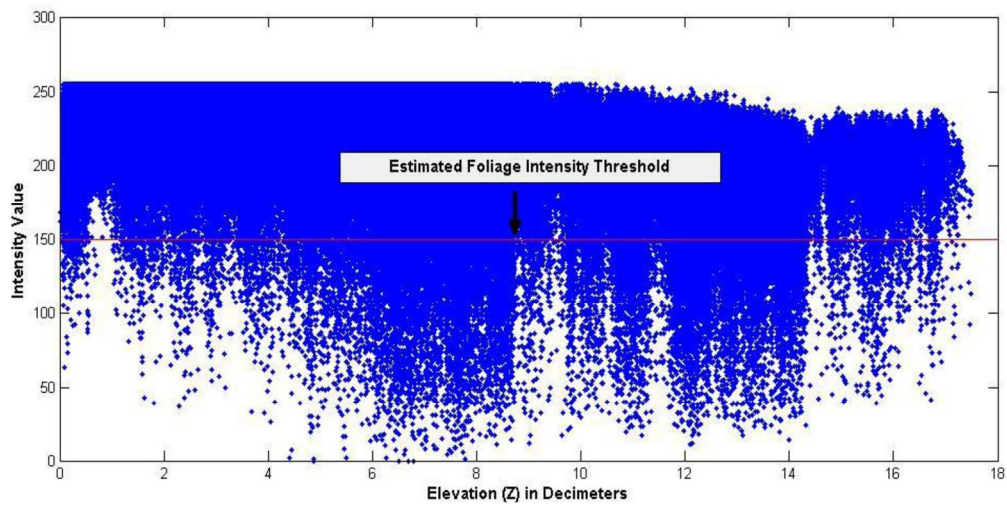


Birds eye view of detection slice taken at 1.37m above the lowest point in the point cloud. Unrestricted subset included branches and foliage located within the height range of the slice.
769x552mm (72 x 72 DPI)



Birds eye view of detection slice taken at 1.37m above the lowest point in the point cloud. Unrestricted subset included branches and foliage located within the height range of the slice.

778x542mm (72 x 72 DPI)



Graph of intensity values versus elevation for 0.1m height bins. Subjectively determined subset threshold is shown in red.

744x371mm (72 x 72 DPI)